



available at www.sciencedirect.com



journal homepage: www.elsevier.com/locate/biocon



2 **Burning biodiversity: Woody biomass use by commercial**
 3 **and subsistence groups in western Uganda's forests**

4 **Lisa Naughton-Treves^{a,b,*}, Daniel M. Kammen^c, Colin Chapman^{d,e}**

5 ^aDepartment of Geography, 550 N. Park Street, University of Wisconsin, Madison, WI 53706, USA

6 ^bCenter for Applied Biodiversity Science, Conservation International, 1919 M St., Washington, DC 20036, USA

7 ^cEnergy and Resources Group, Goldman School of Public Policy, 310 Barrows Hall #3050, University of California,
 8 Berkeley, CA 94720-3050, USA

9 ^dDepartment of Anthropology and McGill School of Environment, 855 Sherbrooke St West, McGill University, Montreal, Canada H3A 2T7

10 ^eWildlife Conservation Society, 2300 Southern Boulevard, Bronx, NY 10460, USA

12

14 **ARTICLE INFO**

15 **Article history:**

16 Received 15 June 2005

17 Received in revised form

18 14 March 2006

19 Accepted 8 August 2006

21 **Keywords:**

22 Deforestation

23 Fuelwood

24 Charcoal

25 Devolution

26 Community-based conservation

ABSTRACT

Woodfuels are the most heavily used energy source in sub-Saharan Africa. We analyzed the ecological impacts and modes of access of five user groups (domestic consumers, gin distillers, brick manufacturers, charcoal producers, and tea companies) drawing biomass energy from natural forests in western Uganda. While domestic consumers use the most species for fuelwood (>50), their consumption is likely sustainable because they generally harvest fast-growing species from fallows on their own land or their neighbors'. Charcoal producers prefer old-growth hardwood species and are responsible for the greatest loss of natural forests. They access forests by finding landholders who, either willingly or through coercion, allow trees on their lands to be cleared. The impact of charcoal production is exacerbated by a license system that undervalues natural forests and rewards rapid harvests across large areas. The tea industry consumes mainly eucalyptus wood (*Eucalyptus* spp.) from corporate plantations, but they indirectly create pressure on natural forests by hiring immigrants who subsequently settle in and clear forest remnants. If such practices continue, forest remnants will soon be exhausted, leaving Kibale National Park as the last natural forest in the region. Forest remnants are a vital source of water, medicinal plants, and energy for local citizens and to protect them from over-exploitation, policy makers should target the charcoal and tea industry for reform. Support for local land management institutions governing access to fallows and successional forests will inevitably enhance the policy interventions.

© 2006 Published by Elsevier Ltd.

48

49 **1. Introduction**

51 **1.1. Woodfuel consumption and forest ecosystems**

52 Approximately half of the wood cut annually worldwide is
 53 used as fuel, and of this amount, nearly 90% is produced

and consumed in developing countries, where firewood and charcoal constitute the primary source of energy for the poor (Okello et al., 2001; Parikka, 2004; Dovie et al., 2004). This reliance is even more pronounced in sub-Saharan Africa, where woodfuels are the dominant energy source, both in terms of primary energy supply and the number of people relying on

* Corresponding author: Address: Department of Geography, 550 N. Park Street, University of Wisconsin, Madison, WI 53706, USA. Tel.: +1 608 262 4846; fax: +1 608 265 3991.

E-mail address: Naughton@geography.wisc.edu (L. Naughton-Treves).

0006-3207/\$ - see front matter © 2006 Published by Elsevier Ltd.

doi:10.1016/j.biocon.2006.08.020

60 them (Bailis et al., 2005). Even in countries with large endow-
 61 ments of fossil fuels like Gabon and Nigeria, woodfuels are a
 62 significant energy source. Heavy reliance on woodfuels can
 63 result in a range of negative environmental impacts, with
 64 both local and global consequences, including forest loss
 65 and degradation, health problems for charcoal producers
 66 and households where biofuels are combusted (Ezzati, 2002),
 67 and increased greenhouse gas emissions (Bailis et al., 2005).
 68 In addition, the expenses associated with both collected and
 69 purchased woodfuel are a major economic burden on the
 70 poor (Ezzati et al., 2001). The African humid tropics may not
 71 yet suffer the acute fuelwood shortages common in arid Afri-
 72 can regions, but fuelwood harvesting in these regions is
 73 extensive and significantly impacts biodiversity.

74 It is estimated that the extraction of wood from tropical
 75 forest for timber, charcoal burning and fuelwood constitute
 76 68% of the proximate causes of deforestation in Africa, 89%
 77 in Asia and 51% in Latin America (Geist and Lambin, 2001).
 78 Even if forest is not entirely cleared, selective harvesting
 79 may change forest composition and ecosystem function
 80 (see Ndangalasi et al., this volume). For example, the har-
 81 vesting of trees from old-growth forest may result in slow
 82 growing species being replaced with faster growing second-
 83 ary species. Secondary growth is more susceptible to fire. If
 84 the burn frequency is sufficiently high, these disturbed for-
 85 ested areas, which were originally mature forest, can be con-
 86 verted to grassland that can be maintained indefinitely by
 87 fire. Unlike other tropical regions, many African mid-eleva-
 88 tion secondary growth species do not provide food for frugiv-
 89 orous birds or primates (Struhsaker, 1997). Thus the
 90 replacement of old-growth trees with secondary growth re-
 91 duces populations of many frugivores (Struhsaker, 1997).
 92 This reduction has been documented to last for over 30 years
 93 (Chapman et al., 2000) and is speculated to last much longer
 94 (Chapman and Chapman, 2004). Depending on the extent of
 95 disturbance produced through harvesting, plant evolutionary
 96 characteristics such as timing of reproduction, resource allo-
 97 cation, reproductive value, seed size, seed crop size, and seed
 98 germination may be significantly altered (Silvertown and
 99 Lovett-Doust, 1993). This may lead to the disruption of eco-
 100 system, community or population structure and changes in
 101 resource and substrate availability or the physical environ-
 102 ment. Changes in ecosystem characteristics such as species
 103 diversity, nutrient output, and biomass as well as changes
 104 that reset succession in one or more sites may be signifi-
 105 cantly disturbed (Mooney and Godron, 1983; White and Pick-
 106 ett, 1985; Turner et al., 1993; Burrows, 1993). In this paper, we
 107 measure patterns of fuelwood extraction in the moist, ever-
 108 green forests of western Uganda and discuss its effect on
 109 biological diversity.

110 1.2. Woodfuel consumption in Uganda

111 The forests of western Uganda contain biological diversity of
 112 global importance (Struhsaker, 1987) and are vital to local
 113 populations for sustaining ecosystem processes and provid-
 114 ing multiple resources, especially fuelwood (Chapman and
 115 Chapman, 1996). Yet rapid population growth (3.3% per year),
 116 expanded commercial charcoal and brick production, as well
 117 as urban and industrial fuelwood demands are fundamen-

118 tally altering the relationship between forests and forest
 119 users with resultant negative impacts on the forest resource
 120 base. In addition, the demand for forest products has intensi-
 121 fied in the context of insecure property rights. The resulting
 122 rapid deforestation in Uganda (~600 km² per year) is threat-
 123 ening the long-term sustainability of land use in the region
 124 (Banana and Gombya-Ssembajwe, 1996).

125 Given Uganda's poor record of state-centered forest con-
 126 servation policies, contemporary researchers and practitio-
 127 ners alike are calling for new approaches to forest
 128 management (Banana et al., 2004). Community-based forest
 129 management has become a popular model, bolstered by polit-
 130 ical and ethical arguments regarding local resource access
 131 and self-determination (Agrawal and Gibson, 1999). At the
 132 heart of the community-based approach is the question of re-
 133 source access—who has the ability and the right to use and
 134 benefit from forest resources? Throughout Uganda, commu-
 135 nity-based forest management is promoted with a vague con-
 136 cept of who constitutes “the local community” (Mugisha,
 137 2002). To evaluate the feasibility of community-based man-
 138 agement, local forest conditions and the ecological impacts
 139 of various user groups need to be thoroughly studied and
 140 understood in a holistic resource availability and use frame-
 141 work. Customary and formal forest access rules must also
 142 be accounted for, as well as the influence of national policies
 143 on local forest management.

144 In this article, we analyze the ecological impacts and
 145 modes of access by five user groups drawing biomass energy
 146 from natural forests in western Uganda. The five groups are
 147 (1) domestic consumers, (2) gin still operators, (3) brick man-
 148 ufacturers, (4) charcoal producers, and (5) tea companies. We
 149 selected these groups because they are embedded differently
 150 in markets and regulatory systems. Each has varying ability to
 151 benefit from forest resources by means of transport, labor,
 152 capital, and political power. Together, these oft-reinforcing
 153 factors shape fuelwood consumption patterns and the result-
 154 ing impacts on ecosystems. They also influence how each
 155 user group responds to growing resource scarcity. By identify-
 156 ing species preferences of these groups, the volumes and
 157 rates of fuelwood harvest, and source forests, we aim to iden-
 158 tify activities that most immediately threaten biodiversity.
 159 Thus we hope to guide efforts to promote sustainable com-
 160 munity-based forest management in western Uganda and
 161 other African forests that are under similar pressure to supply
 162 fuelwood.

2. Study area 163

2.1. Ecological characteristics 164

165 The study area is located around Kibale National Park
 166 (795 km²), in the Kabarole District in western Uganda, lying
 167 immediately northeast of the Rwenzori Mountains. Kibale
 168 National Park holds the last substantial tract of premontane
 169 forest in East Africa (Chapman and Chapman, 1996) (Fig. 1).
 170 Surrounding Kibale National Park is a mosaic of grasslands,
 171 smallholder agriculture, papyrus swamps, tea, eucalyptus
 172 plantations, and patches of natural forests. These forest
 173 patches average 32 ha in size (range 3 to 350 ha) and are lo-
 174 cated almost entirely in wet lowlands or steep slopes.

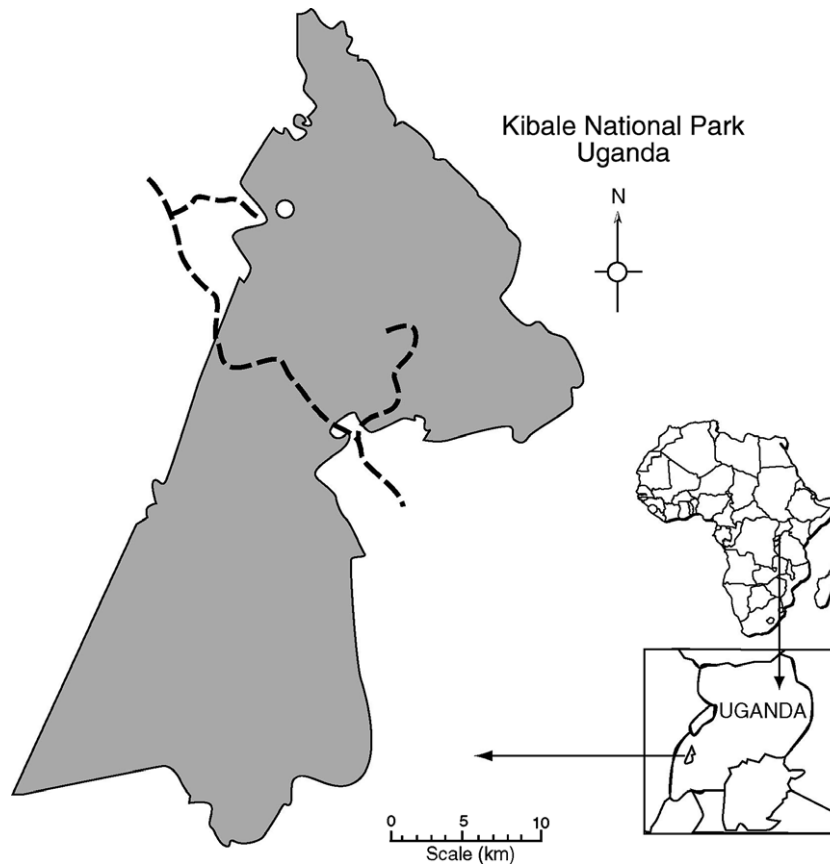


Fig. 1 – Map of study area, Kibale National Park in western Uganda.

175 Foresters have classified the forest in this region as a *Par-*
 176 *inari* forest, distinguished on photo aspect maps by large
 177 spreading crowns of *Parinari excelsa* (Skorupa, 1988; Kings-
 178 ton, 1967). At this elevation (1370–1525 m), the presence of
 179 *P. excelsa* and the subdominants (*Pouteria altissima*, *Olea cap-*
 180 *ensis*, *Newtonia buchananii*, and *Chrysophyllum gorungosanum*)
 181 is associated with old-growth forest (Osmaston, 1959). When
 182 discussing specific tree species used as fuelwood, we indi-
 183 cate if they are old-growth, mid-successional, early succes-
 184 sional, or exotic species following Zanne and Chapman (in
 185 press) and Hamilton (1991). As is typical of many tropical
 186 tree communities, tree growth rates in the region are highly
 187 variable among species (Chapman and Chapman, 2004).
 188 Species typically found in old-growth or mature trees have
 189 growth rates of between 1 and 3 cm diameter at breast
 190 height (DBH) per year, while species colonizing gaps or dis-
 191 turbed areas can have growth rates exceeding 10 cm DBH/
 192 year and can reach >15 m in height and >10 cm DBH in just
 193 5 years (Chapman, unpublished data). While rates of seed
 194 dispersal into areas of disturbed forest are not reduced,
 195 recruitment of seedlings and saplings is very poor, due to
 196 competition with grasses and an aggressive herbaceous
 197 layer (Chapman and Chapman, 1997; Paul et al., 2004). Areas
 198 of mature forest are not typically susceptible to fire, while
 199 areas of degraded forest are (Chapman et al., 1999; Lwanga,
 200 2003).

2.2. Social context and rules of forest access

201

The dominant ethnic group in the area is the Batoro people. 202
 Since their arrival in Kabarole District¹ during the 19th cen- 203
 tury (Naughton-Treves, 1999), the Batoro have developed a lo- 204
 cal system of ownership and forest use, incorporating both 205
 spatially explicit resource domains (e.g., royal and village for- 206
 ests) and user rights to specific tree species (Kaipiriri, 1997). 207
 Royal forests are managed similarly to village forests except 208
 that a special tax is levied on any commercial users of the for- 209
 est and is collected by a representative of the *Omukama* (king) 210
 (Kasenene, J., pers. comm.). As is common in East Africa, for- 211
 est access rules are complex and include overlapping tenure 212
 claims (Rocheleau and Edmunds, 1997). Most forest patches 213
 and swamp forests are considered village property that is sub- 214
 divided into individually managed parcels. Few individuals 215
 have legal title to their land; they claim it instead under cus- 216
 tomary rules. Some elements of communal ownership per- 217
 sist. Individual owners are typically obligated to give kin 218
 and neighbors permission to use natural forests and old fal- 219
 lows for cooking fuelwood, medicinals, drinking water, and 220
 other subsistence purposes. 221

¹ Prior to independence, the region was known as the Toro Kingdom.

222 Traditional forest property regimes were undermined by
 223 state-imposed regulations and commercial timber markets
 224 during and immediately after the colonial period. During this
 225 time, the Ugandan Forest Department assumed ownership of
 226 all large blocks of forests (including Kibale Forest in 1932) and
 227 managed them for timber extraction. Local people were
 228 prohibited from harvesting resources in state forests and
 229 were expected to rely instead on surrounding forest patches.
 230 Following independence, between 1971 and 1986, Uganda
 231 experienced war, severe economic recession, and the disinte-
 232 gration of the state. The Forest Department lost control of
 233 Kibale and other forest reserves (Hamilton, 1984), and during
 234 this turbulent period, the population density in Kabarole Dis-
 235 trict tripled (from 27 to 97 people per km²) due to high fertility
 236 rates and the immigration of tens of thousands of Bakiga peo-
 237 ple from southwestern Uganda (World Bank, 1993). In 1990,
 238 the post-war Ugandan government “upgraded” Kibale from
 239 a reserve to a National Park and used force to control illicit
 240 use of park resources (Feeny, 1998). More recently, park man-
 241 agers and local leaders have begun to discuss collaborative
 242 management at Kibale, and to allow some neighboring com-
 243 munities to use non-timber forest products within specified
 244 zones (KNP General Management Plan 2005). Outside the
 245 park, the continued rapid population growth (3.4%, among
 246 the fastest in Uganda) and high demand for fuelwood and
 247 charcoal has intensified pressure on forests (Government of
 248 Uganda, 2002).

249 2.3. Current fuelwood use and patterns of deforestation

250 Over 95% of Kabarole’s people rely exclusively on wood for en-
 251 ergy (Government of Uganda, 2002). Charcoal production for
 252 regional and national urban markets is expanding (Chapman
 253 and Chapman, 1996). As elsewhere in East Africa, charcoal is
 254 produced by burning logs in oxygen-restricted pyrolytic con-
 255 ditions inside earth kilns (known locally as ‘heaps’) averaging
 256 5.7 m³ in size (range 1.7–10.9 m³). Brick production is also
 257 growing to meet construction demands of growing urban
 258 and semi-urban areas. The commercial production of banana
 259 beer and gin also requires fuelwood resources (~60% of small-
 260 holder land is cultivated with brewing bananas) (Naughton-
 261 Treves, 1998). The area devoted to tea cultivation in Kabarole
 262 has expanded by 2000 ha (~10%) in the past 40 years (Mulley
 263 and Unruh, 2004), much of this within 10 km of Kibale Park,
 264 where tea production expanded six fold between 1955 and
 265 1988 (Mugisha, S., unpublished data). Mulley and Unruh
 266 (2004) explain that this expansion intensifies pressure on for-
 267 ests in two ways: first, tea companies require significant
 268 amounts of eucalyptus to dry their tea (roughly one hectare
 269 of eucalyptus is needed per every three hectares of tea), and
 270 second, they import laborers from outside Kabarole,² many
 271 of whom eventually leave the tea companies and establish
 272 homesteads on land near the park boundary.

273 Amidst growing demands for forest resources and uncer-
 274 tain access rules, deforestation in Kabarole has continued
 275 apace. According to remote sensing analysis by Mulley and

Unruh (2004), between 1955 and 2001 forest declined by 276
 7967 ha outside of Kibale National Park, while increasing by 277
 10,823 ha within the Park, due to forest regrowth in formerly 278
 cultivated areas. A similar analysis by Naughton-Treves 279
 et al. (unpubl.) along the western boundary of Kibale showed 280
 that forest loss inside the park within 500 m of the boundary 281
 proceeded at 0.2% per year between 1995 and 2001. During 282
 this same period, forest declined outside the park by 2.3– 283
 3.8% per year, with the fastest rates occurring within 1 km 284
 of the park boundary. This peak in deforestation rates near 285
 the park was similarly observed for other forest reserves of 286
 the Albertine Rift (Plumptre, 2002, and see Plumptre et al., this 287
 volume). The accelerated deforestation near the park bound- 288
 ary is cause for concern given the deleterious effects of isola- 289
 tion on biodiversity conservation (Brooks et al., 1999). As in 290
 the case of other African forest parks, the pattern of defores- 291
 tation also portends heightened pressure on Kibale in the fu- 292
 ture if surrounding forests are exhausted (Struhsaker et al., 293
 2005). 294

3. Methods 295

We focused our monitoring and interviews along the western 296
 side of Kibale National Park, up to 10 km from the boundary 297
 (Fig. 1). Due to our long history (15+ years) of working in this 298
 area and our basic knowledge of the Toro language, citizens 299
 were generally comfortable discussing their fuelwood uses 300
 and opinions about forest management with us. We also as- 301
 sured respondents we would protect their anonymity. Expert 302
 help from local resident field assistants (both Batoro and Bak- 303
 iga), further improved the data quality. Despite these advan- 304
 tages, illicit uses were likely underestimated by some 305
 respondents. 306

3.1. Survey 307

To obtain a rough estimate of the proportion of the local res- 308
 idents that belonged to different fuelwood user groups, we 309
 surveyed people from 160 households divided into three 310
 groups: those located adjacent to a natural forest patch, adja- 311
 cent to Kibale National Park, or in the second or third row of 312
 properties away from any forest edge. This sampling design 313
 was sensitive to the strongly spatial underpinnings of tradi- 314
 tional forest access rules. We asked each respondent if any- 315
 one in the household had engaged in cooking, producing 316
 gin, or brick making during the past year. We also noted if 317
 they had a woodlot on their property. 318

3.2. Monitoring fuelwood use 319

Between 2000 and 2001, using a participant observer ap- 320
 proach, we monitored 86 fuelwood combustion “episodes”, 321
 including 32 domestic cookers, 18 brick makers, 26 charcoal 322
 producers, and 24 banana gin distillers. We used a stratified 323
 semi-random method to select individuals lying along a con- 324
 tinuum of increasing distance from Kibale. This design was 325
 easy to follow in the case of domestic cookers who are pres- 326
 ent throughout the landscape (Table 1). Sampling the other 327
 fuelwood user groups was more difficult given their non-ran- 328
 dom distribution on the landscape. For example, only one still 329

² Roughly half of the local tea labor force is recruited from areas beyond Kabarole District (Lameck, K., Manager, J. Finlay Tea Ltd., pers. comm., 2006).

Table 1 – A description of the sample of informants used to describe patterns of fuelwood use in the region adjacent to Kibale National Park, Uganda

User group	Total sample size	% Women sampled (as % of people actively involved)	No. of samples			Average duration of each combustion episode (i.e., period monitored, in hours)
			<1 km from Kibale National Park	<1 km from other natural forest	Kibale or other natural forest	
Domestic consumers	28	96	7	13	8	4
Brick makers	22	7	0	6	16	44
Gin distillers	22	0	13	8	1	13
Charcoal producers	26	5	12	14	0	128

330 and no charcoal heaps were found >500 m from the edge of
 331 Kibale or another patch of natural forest. Conversely, all but
 332 two brick kilns were located within 100 m of a major road,
 333 and thus >5 km from the park.

334 In each case, we identified and weighed all logs and wood
 335 pieces designated for use in a particular combustion episode.
 336 Logs <40 kg were weighed (wet weight). The mass of larger
 337 logs was estimated from measuring the volume (length \times πr^2)
 338 and multiplying this by the wood density for the 12 frequently
 339 used species: *Vernonia* spp., *Acanthus pubescens*, *Maesa lanceo-*
 340 *lata*, *P. excelsa*, *Bridelia micrantha*, *Macaranga schweinfurthii*,
 341 *Eucalyptus* spp., *Olea welwitschii*, *Diospyros abyssinica*, *Fagaropsis*
 342 *angolensis*, *N. buchananii* and *Prunus africana*. This resulted in
 343 more precise estimates for smaller pieces than for large logs.
 344 Wood density for these 12 species was calculated by volumetric
 345 displacement of water by fresh wood pieces. For other species,
 346 we used wood density listed in Brown (1997). We
 347 identified the species of wood by visual observation. If we
 348 did not recognize the species, the fuelwood users assisted in
 349 its identification. The final proportion of “unknown” species
 350 varied among fuelwood uses. Seven percent of fuelwood
 351 pieces used in domestic cooking could not be identified, 7%
 352 of pieces used to make charcoal were also not identified, 2%
 353 remained unidentified for stills, and <1% for brick kilns. These
 354 field methods were inappropriate for measuring fuelwood use
 355 by tea processing factories, given the industrial scale of their
 356 operation. Instead we relied on the records of the single tea
 357 processing plant in the study area (Kiko processing plant of
 358 Rwenzori Highland Tea Company, later purchased by James
 359 Finlays (U) Ltd.).

360 3.3. Semi-structured interviews

361 We took advantage of the many hours spent monitoring fuel-
 362 wood use to conduct in-dept, semi-structured interviews with
 363 individuals from each of the user groups. We asked people
 364 where they had obtained each piece of fuelwood, if/how they
 365 negotiated access to the fuelwood, if they perceived any
 366 shortage of fuelwood, and if so, who or what they thought
 367 was to blame for the shortage. We also tallied outputs of
 368 the combustion episode (e.g., volume of gin or charcoal pro-
 369 duced, number of bricks, number of meals cooked). To obtain
 370 a broader picture of forest access rules, we interviewed 11 key
 371 informants, including four local council chairmen, two Forest
 372 Department officials in the regional municipal office (Fort Por-
 373 tal), three charcoal transporters, and two small-scale charcoal
 374 vendors.

375 3.4. Data analysis

376 Descriptive statistics was the major source of analysis. How-
 377 ever, non-parametric statistics (e.g., Kruskal–Wallis and
 378 Mann–Whitney *U*-Test) were used under various regimes.
 379 The Kruskal–Wallis test was used to compare the differences
 380 in the amount of wood utilized by the five groups as well as
 381 the number of species used by the user groups in each com-
 382 bustion episode. The Mann–Whitney *U* test was to ascertain
 383 the differences in the mean biomass consumption between
 384 the user groups. We used the CITES classification to identify
 385 the threat status of species.

386 4. Results

387 The composition of the 160 respondents in our field survey
 388 roughly accords with the make-up of Kabarole's general pop-
 389 ulation. There were 48% women and 52% men, and 29% Kiga,
 390 68% Toro and 3% ‘Other ethnicity’ (compared to 28% Kiga, 70%
 391 Toro and 2% ‘Other’ according to the 2002 Ugandan govern-
 392 ment census of Kabarole).

393 4.1. Physical patterns of use and sustainability

394 The results of the survey of 160 residents reveal that all respon-
 395 dents rely on firewood for cooking, no matter where they live
 396 (Table 2). Charcoal production was the second most frequent
 397 use (17.8%), but this activity was confined largely to natural for-
 398 est edges and was especially prevalent around village-man-
 399 aged forest patches. Those who distilled banana gin resided
 400 mainly along the edge of forest remnants and Kibale Park and
 401 comprised 14.4% of the total respondents. Only 2.5% of respon-
 402 dents baked bricks. Overall the amount of biomass harvested
 403 per combustion episode differed significantly amongst the dif-
 404 ferent users (Kruskall–Wallis = 68.01, $P < 0.001$), with charcoal
 405 using more biomass than any other group (Mann–Whitney
 406 tests between pairs $P < 0.02$ or less). Brick making used more
 407 woody biomass than stills ($P = 0.003$) or women collecting cook-
 408 ing firewood ($P < 0.001$), and biomass collection for stills being
 409 more than collection for cooking firewood ($P < 0.001$).

410 As a group, women gathering firewood for cooking
 411 (‘domestic consumers’) used the greatest number of woody
 412 species (50) (Table 3). At the other extreme was the tea pro-
 413 cessing plant, which relied entirely on one species of eucalypt-
 414 us (*Eucalyptus grandis*) to fuel its tea leaf driers. Brick makers,
 415 gin distillers, and charcoal producers all used a comparable
 416 number of species (~26). The number of woody species used

Table 2 – The results of the survey of fuelwood use by residents (N = 160) near Kibale National Park, Uganda

Households' location	N	% Ethnic group ^a	% Women interviewed	During past year, % using fuelwood			
				Cook	Distill gin	Bake bricks	Produce charcoal
Household adjacent to Kibale National Park	52	35 K 62 T 3 O	42	100	26.9	1.9	9.6
Household adjacent to forest patch	56	18 K 77 T 5 O	43	100	16	4	21
Household in second or third 'row' from forest	51	35.3 K 60.7 T 4.0 O	58.8	100	7.8	0	2.0
Total	160	28.8 K 66.2 T 5.0 O	48.1	100	14.4	2.5	17.8

a T = Batoro people, K = Bakiga people, O = other ethnicity (includes people of Nyoro, Ankole, Hutu and other ethnic heritages).

Table 3 – A description of the woody species used and annual combustion of five groups of fuelwood users adjacent to Kibale National Park, Uganda

Fuelwood user (n)	Total no. of spp used	Top species (by % total weight for those accounting for >2% and successional stage ^a)	Annual combustion ^b (tons/yr/user)
Domestic consumer (28 households)	50	<i>Vernonia</i> spp. (14.8%) – ES <i>Acanthus pubescens</i> (10.9%) – ES <i>Maesa lanceolata</i> (6.5%) – ES <i>Parinari excelsa</i> (4.3%) – OG <i>Bridelia micrantha</i> (3.9%) – ES <i>Psidium guavaja</i> (3.9%) – EX <i>Ficus brachylepis</i> (3.2%) – ES <i>Sesbania sesban</i> (3.2%) – ES	4.3
Brick maker (18 kilns)	22	<i>Eucalyptus</i> spp. (78.4%) – EX <i>Persea americana</i> (5.6%) – EX <i>Sapium ellipticum</i> (4.4%) – ES	96.4
Gin distiller (22 stills)	27	<i>Prunus africana</i> (33.2%) – ES <i>Parinari excelsa</i> (11.1%) – OG <i>Macaranga scheiwinfurthii</i> (4.8%) – ES	6.1
<i>Olea welwitschii</i> (2.3%) – MS			
Charcoal producer (26 heaps)	26	<i>Diospyros abyssinica</i> (18.1%) – MS <i>Olea welwitschii</i> (12.0%) – MS <i>Fagaropsis angolensis</i> (10.3%) – MS <i>Newtonia buchananii</i> (7.8%) – OG <i>Parinari excelsa</i> (6.1%) – OG	50.8
Tea factory (1)	1	<i>Eucalyptus</i> spp. (100%) – EX	2535–3000

a Successional status of species as per Zanne et al. (in press) with OG: old growth, MS: mid-successional, ES: early successional, EX: exotic.

b Calculated by multiplying biomass used per burning episode by number of burning episodes every year, as estimated by informants.

417 during each combustion episode also differed significantly
418 among the five user groups (Kruskall–Wallis = 9.264, $P =$
419 0.026). The number of species burned during an average
420 brick-making project was less than that taken to fuel stills
421 (pairwise comparison using Mann–Whitney, $Z = 3.1$, $P =$
422 0.008), produce charcoal ($Z = 3.3$, $P = 0.002$), and cook food

($Z = 3.4$, $P < 0.002$). Fueling stills typically involved a species
423 harvest that was comparable to charcoal production and used
424 marginally more species than the average used during a day's
425 cooking ($Z = 1.98$, $P = 0.054$). Finally, the number of species
426 used in charcoal production was similar to that collected by
427 women for fuelwood. The major difference seems to be that
428

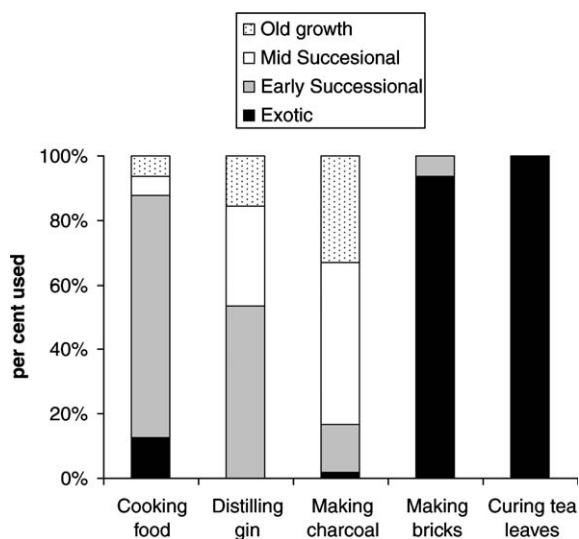


Fig. 2 – Type of woody species used in Kabarole, Uganda.

in producing bricks people are more selective and only use trees found near roads (e.g., *Eucalyptus* spp. and other exotic species).

More important than a simple tally of the average number of species used per combustion episode is the type of species harvested by the different groups (Fig. 2). Women relied mainly on fast-growing early successional species like *Vernonia* spp. for cooking food (Table 3). In previous research (Naughton-Treves and Chapman, 2002), we calculated that each household in the study area would require roughly 0.5 ha of land fallowed for ~4 years to meet their fuelwood needs for cooking (8.4 kg per day). Brick makers meanwhile primarily harvested eucalyptus trees. By contrast, gin distillers and charcoal producers burned slow-growing hardwood species such as *P. excelsa*, *N. buchananii*, and *O. welwitschii*. These species are rapidly disappearing from forests outside the park and provide important food resources for frugivores in the region. During interviews, respondents ranked these three old-growth species as “most scarce” along with two early successional species: *B. micrantha* and *P. africana*. *P. africana* is listed in Appendix II of CITES and is highly valued for its medicinal properties (Anonymous, 2005; Fashing, 2004).

4.2. Forms of acquiring access to fuelwood and coping with scarcity

Table 4 reveals the distinct pattern of extraction for each user group. On average, women collect just over half their firewood for cooking from fallow land and woodlots on their own property. Their second major source is woody species growing on their neighbors’ land. During interviews, women explained that this is customary and that it would be rude for their neighbor to refuse them “small sticks” for cooking. Women also collected fallen branches of hardwoods from forest remnants, especially when they needed to prepare slow-cooking food. Average time spent searching for and collecting firewood was 1.1 hours/day (range 0.5–3), a relatively low value compared to the travel and collection times of nearly 5 h/day recorded in some other parts of Africa (Kammen, 1995).

The brick makers relied primarily on eucalyptus trees grown on neighbors’ woodlots. They obtained this wood by bargaining with their neighbors and offering them a percentage of the profit from the sale of the bricks. Due to the significant cost of transporting bricks, brick makers located their kilns near major roads and thus they were too far from natural forest remnants to draw on them. Transport costs apparently did not constrain other fuelwood extractive activities, such as gin distillation and charcoal manufacturing.

Gin distillation is typically a group project in which various individuals contribute banana mash, distillation equipment (barrel and pipes), and labor. Distillers pay 15,000/= (~US \$8.1 in 2003) each year to the subcounty tenderer for the right to produce gin. Distillation is typically carried out where banana fields abut natural forest, and where a small stream flows. Small fires need to be maintained continuously for 12 h, thus slow-burning hardwood logs are favored. These are mainly drawn from fallen branches or logs within an area <500 m of the still, including within the Park occasionally. If logs are drawn from forest belonging to a neighbor, that neighbor receives a portion of gin sale profits.

Charcoal production is officially regulated by a license system where individuals pay 26,800 per month (equivalent to ~US \$14.60, comprised of a \$8.70 transport fee, \$5.88 burn fee) to produce as much as they can from anywhere in the District. There is an active, illicit trade in sharing and duplicating these licenses. Individuals living in remote forested areas do not buy licenses; rather, it is the intermediary who buys and transports the charcoal to town must have the li-

Table 4 – Source of fuelwood by land property type for four use groups

Use groups	Total kg traced to source	Own property		Neighbors’ property		Village forest patch ^c	Kibale National Park
		Fallow ^a	Woodlot ^b	Fallow ^a	Woodlot ^b		
Domestic cookers	483.2	38.2%	18.5%	23.4%	4.8%	13.1%	1.6%
Gin distillers	2829.3	5.8%	2.6%	2.0%	0%	80.5%	9.1%
Charcoal manufacturers	28190.8	0	0	7.3%	0%	92.3%	0.4%
Brick makers	19320.6	3.9%	12.4%	16.0%	67.2%	0.2%	0

a “Fallow” is a broad category including naturally regenerating fallows 4–20 years old as well as areas of brush managed amidst sparsely planted banana fields.

b “Woodlot” refers to planted areas of exotic trees, primarily *Eucalyptus grandis*.

c Includes royal patches. See Naughton-Treves (1999) for a history of royal claims to wildlife and forests in Kabarole.

cense. Certain individuals (often Kiga immigrants) specialize in manufacturing charcoal, an arduous job commonly referred to as “poor man’s work”. The access rules for charcoal producers are often unclear. In fact, during several interviews, landowners asked the first author to explain who had the right to produce charcoal. Most typically, landowners who agree to have charcoal produced from a portion of their territory of village forest receive one or two sacks of charcoal (each worth ~2500/= or ~\$1.40 US) as payment (average yield per “heap” is 17 sacks). Individuals residing on Royal forested land officially have no right to demand payment or to deny permission to charcoal producers. In such cases, the King’s representative collects one or two sacks of charcoal per heap. Some individuals, single women in particular, complained that they could not refuse “men who come with papers” (i.e., licenses). Evidently, customary norms of access do not govern the charcoal business, but neither do legal codes. For example, it is illegal to clear forest alongside streams, yet the majority of charcoal production episodes occurred close to water given that this is where the last natural forest grows.

Unlike the other fuelwood uses described so far, tea is part of the formal national and international economy. Tea processing factories in Kabarole appear to abide by environmental laws. Most keep careful records of their fuelwood use. Some publicly post-rules for environmental stewardship, and one company, J. Findlay (U), attained accreditation for Environmental Management under the ISO 14001 system. The tea factory in the study area now burns only eucalyptus trees grown on their private property (Table 4). But some other factories continue to buy eucalyptus from local farmers. Moreover, as tea production continues to expand in Kabarole, more land for raising eucalyptus is required, which the tea companies acquire by buying land from local people. Under this practice, eucalyptus is often planted in wetlands, although this latter practice is illegal (Mulley and Unruh, 2004).

5. Discussion

5.1. Proximate concerns for loss of biological diversity

The patterns of fuelwood extraction we observed in western Uganda accord with studies elsewhere in tropical Africa (Parrotta et al., 2002). Woodfuels continue to provide nearly all the energy required for cooking and heating in rural and urban households in Uganda (Ugandan National Forestry Authority Report, 2004, cited in {Zusammenarbeit, 2005 #1213}). The high household demand for fuelwood and charcoal coupled industrial demands (including expanded cultivation of tea) is driving rapid deforestation. Using old growth, hardwood species as fuel is not sustainable, particularly for the manufacture of charcoal, given the large volumes of wood consumed. The regeneration of hardwoods on fallow land is very slow (Chapman et al., 1999; Naughton-Treves, 2002). For example, to perpetually supply a charcoal manufacturer with 50 tons of native hardwoods per year would require an extremely large area—depending on regeneration and growth rates possibly as much as 1000 km² of land area. This represents the standard accumulation of native hardwoods on fallow land managed by smallholders. If lands were to be managed specifically for hardwood production (e.g., liberation thinning,

enrichment planting) biomass accumulation would be substantially higher. At present, however, local citizens cannot afford to manage land for such long time horizons, or they are too uncertain of their long-term ownership of forests. Thus, even though home cooking consumers outnumber charcoal producers by >5–1, it is the charcoal producers’ activities that immediately threaten natural forests.

Meanwhile, women’s use of firewood around Kibale appears to be sustainable due to their reliance on early successional species from fallow lands. However, the campaign to privatize land in Uganda could potentially undermine women’s access to firewood given that they now rely on customary access rules to use wood from one another’s fallows and fields (Rocheleau and Edmunds, 1997). Supporting these customary access rules is in the best interest of forests and local communities.

At the other extreme, the tea industry’s impact on forests is shaped by formalized national and international regulations. Some companies are attempting to produce tea according to international fair-trade criteria, which include protecting natural forests. While this is a laudable goal, the tea industry will still indirectly threaten Uganda’s western forests as long as companies buy up fallow land for eucalyptus plantations and import large numbers of workers for their plantations and processing plants.

5.2. Policy and management implications

As in many tropical regions (Parrotta et al., 2002), much of western Uganda’s natural forests are not being “managed” so much as “mined” for fuelwood (charcoal in particular). Similar patterns have been observed in neighboring Tanzania, where commercial harvesting for charcoal overrides ecological impacts from all other harvesting purposes because of powerful economic incentives and the wide range of species and size classes used to produce charcoal (Luoga et al., 2002). In western Uganda, as elsewhere in Africa, local citizens obtain negligible income from charcoal production. Instead, profits accrue to charcoal transporters and wholesale traders, as well as to those engaged in the illicit “license” trade (Brouwer and Falcao, 2004).

Policy reform aimed at improving the sustainability of charcoal manufacture is urgently needed to slow biodiversity loss and protect ecological services and local livelihoods. Such reform may be complex due to the fact that while charcoal has negative environmental impacts, using charcoal instead of firewood has positive consequences for human health. Compared to wood, charcoal use contributes much more to biodiversity loss and climate change (Bailis et al., 2005). On the other hand, charcoal burns more cleanly than wood and switching from wood to charcoal can reduce a household’s levels of particulate matter and carbon monoxide by as much as 90% (Ezzati and Kammen, 2002). This in turn can decrease the incidence of respiratory illness in children by 50%. Thus, although choosing charcoal over wood can negatively impact the environment, it can also reduce the incidence of illness in the household.

Management strategies that encourage more sustainable charcoal production, transport, and end-use are only in their infancy. Many experts call for more efficient, modern kilns

611 (Okello et al., 2001). Others argue that marketing and license
 612 rules should be reformed to foster prices that better reflect
 613 charcoal's inherent ecological and social costs (Lew and Kam-
 614 men, in press; Ribot, 1998). Such reform is often politically dif-
 615 ficult due to urban consumers' demand for cheap energy. For
 616 example, Nigeria's efforts to slow fuelwood consumption
 617 have been "unsuccessful" due to the high costs of alternative
 618 fuels (Kersten et al., 1998). These experiences have lead some
 619 analysts to call for greater subsidies for electric power for ur-
 620 ban consumers, as a way to alleviate some of the pressure on
 621 forests {Ministry of Agriculture, 2004 #1214}.

622 5.3. Concluding remarks

623 At present, the greatest threat to Uganda's natural forests is
 624 charcoal production, an economic activity that at present
 625 does not conform to customary forest access rules or codified
 626 legal regulations. Because of abundant, inexpensive access,
 627 Ugandans have one of the highest levels of consumption of
 628 firewood and charcoal in Africa (1.77 m³ fuelwood per person
 629 per year, Brouwer and Falcao, 2004). However, this access
 630 comes at the expense of the country's natural forests and bio-
 631 logical diversity, and the long-term maintenance of local eco-
 632 system services (e.g., watershed and soil protection). Outside
 633 of national parks, old-growth species are becoming increas-
 634 ingly rare because harvest patterns favor short-lived, early
 635 successional tree species or grasslands (Okello et al., 2001).
 636 The Ugandan National Forest Authority projects that the cur-
 637 rent national consumption rate of 20 million tons of wood per
 638 year will triple by 2025 [Zusammenarbeit, 2005 #1213]. Under
 639 such a scenario, old-growth forests will likely be extirpated
 640 outside of parks and reserves. To conserve Uganda's remain-
 641 ing natural forests, reform in the charcoal sector is urgently
 642 needed, as well as coordination of policies across nations
 643 due to illicit wood harvesting. National energy plans ought
 644 to incorporate biodiversity and livelihood concerns.

645 Acknowledgements

646 This project benefited from first-rate fieldwork by Pascal Ba-
 647 guma. The research was approved by the Ugandan National
 648 Science and Research Council and was supported by a US Na-
 649 tional Science Foundation Grant (#98-10144) from the Geogra-
 650 phy and Regional Sciences Division. Colin Chapman's studies
 651 in the Kibale region were funded by the Wildlife Conservation
 652 Society, National Science Foundation (USA), and NSERC (Can-
 653 ada). Daniel Kammen thanks the Energy Foundation for con-
 654 tinued support, and the University of California Class of 1935.
 655 This manuscript was significantly improved by comments
 656 from A. Banana, N. Cordeiro, D. Dovie, J. Kasenene, N. Lind-
 657 eman and P. Witucki.

658 REFERENCES

659 Agrawal, A., Gibson, C., 1999. Enchantment and disenchantment:
 660 the role of community in natural resource conservation. *World*
 661 *Development* 27, 629–649.

Anonymous, 2005. Ancient Medicinal Tree Threatened with
 Extinction; Tree is leading Remedy for Prostate Disorders
 Worldwide, Washington, DC and Nairobi, Kenya (accessed
 15.05.2005). 662–665

Bailis, R., Ezzati, M., Kammen, D., 2005. Mortality and greenhouse
 gas impacts of biomass and petroleum energy futures in
 Africa. *Science* 308, 98–103. 666–668

Banana, A.Y., Gombya-Ssembajjwe, W., 1996. Successful Forest
 Management: The Importance of Security of Tenure and Rule
 Enforcement in Ugandan Forests. International Forestry
 Resources and Institutions Research Program Series,
 Bloomington, Indiana. 669–673

Banana, A.Y., Vogt, N.D., Gombya-Ssembajjwe, W., Bahati, J., 2004.
 In: The Commons in an Age of Global Transition: Challenges,
 Risks and Opportunities, vol. 10, International Association for
 the Study of Common Property, Oaxaca, Mexico. Available
 from: <<http://dlc.dlib.indiana.edu/archive/00001552/>>. 674–678

Brooks, T.M., Pimm, S.L., Oyugi, J.O., 1999. Time lag between
 deforestation and bird extinction in tropical forest fragments.
Conservation Biology 13, 1140–1147. 679–681

Brouwer, R., Falcao, M.P., 2004. Wood fuel consumption in Maputo,
 Mozambique. *Bioenergy* 27, 233–245. 682–683

Brown, S., 1997. Estimating Biomass and Biomass Change of
 Tropical Forests. Food and Agriculture Organization of the
 United Nations, Rome. 684–686

Burrows, C.J., 1993. Processes of Vegetation Change. UNWIN
 HYMAN, London. 687–688

Chapman, C.A., Chapman, L.J., 1996. Mid-elevation forests: a
 history of disturbance and regeneration. In: McClanahan, T.R.,
 Young, T.P. (Eds.), *East African Ecosystems and their*
Conservation. Oxford University Press, New York, pp. 385–400. 689–692

Chapman, C.A., Chapman, L.J., 1997. Forest regeneration in logged
 and unlogged forests of Kibale National Park, Uganda.
Biotropica 29, 396–412. 693–695

Chapman, C.A., Chapman, L.J., 2004. Unfavorable successional
 pathways and the conservation value of logged tropical forest.
Biodiversity and Conservation 13, 2089–2105. 696–698

Chapman, C.A., Chapman, L.J., Kaufman, L., Zanne, A.E., 1999.
 Potential causes of arrested succession in Kibale National
 Park: growth and mortality of seedlings. *African Journal of*
Ecology 37, 81–92. 699–702

Chapman, C.A., Balcomb, S.R., Gillespie, T.R., Skorupa, J.,
 Struhsaker, T.T., 2000. Long-term effects of logging on
 African primate communities: a 28 year comparison from
 Kibale National Park, Uganda. *Conservation Biology* 14,
 207–217. 703–707

Dovie, D., Witkowski, E.T.F., Shackleton, C.M., 2004. The fuelwood
 crisis in southern Africa relating fuelwood use to livelihoods in
 a rural village. *GeoJournal* 60, 123–133. 708–710

Ezzati, M., Kammen, D., 2002. Household energy, indoor air
 pollution and public health: research and policy needs in
 developing countries. *Annual Review of Energy and the*
Environment 27, 1–38. 711–714

Ezzati, M., Singer, B., Kammen, D., 2001. Towards an integrated
 integrated framework for development and economic policy:
 the dynamics of environmental Kuznets curves. *World*
Development 29, 1421–1434. 715–718

Fashing, P.J., 2004. Mortality trends in the African cherry (*Prunus*
africana) and the implications for colobus monkeys (*Colobus*
guereza) in Kakamega Forest, Kenya. *Biological Conservation*
 120, 449–459. 719–722

Feeny, P., 1998. *Accountable Aid*. Oxfam, London. 723

Geist, H.J., Lambin, E.F., 2001. What Drives Deforestation? A Meta-
 analysis and Underlying Causes of Deforestation based on
 Subnational Case Study Evidence. Ciaco Printshop, Louvaine-
 la Neuve. 724–727

Government of Uganda, 2002. 2001 National Housing and Rural
 Settlement Census. Census Bureau, Kampala, Uganda. 728–729

- 730 Hamilton, C.A., 1984. Deforestation in Uganda. Oxford University
731 Press, Nairobi, Kenya. 775
- 732 Hamilton, A., 1991. A Field Guide to Ugandan Forest Trees.
733 Makerere University Press, Kampala, Uganda. 776
- 734 Kaipiriri, M., 1997. Local use of non-timber products. Master's
735 Thesis, Makerere University, Kampala, Uganda. 777
- 736 Kammen, D., 1995. Cookstoves for the developing world. *Scientific
737 American* 273, 72–75. 778
- 738 Kersten, I., Baumbach, G., Oluwole, A.F., Obioh, I.B., Ogunsola,
739 O.J., 1998. Urban and rural fuelwood situation in the
740 tropical rain-forest area of south-west Nigeria. *Energy* 23,
741 887–898. 779
- 742 Kingston, B., 1967. Working plan for the Kibale and Itwara Central
743 Forest Reserves, 2nd ed. Uganda Government Forest
744 Department, Kampala, Uganda. 780
- 745 Lew, D., Kammen, D., in press. Social and environmental impacts
746 of charcoal. *Energy Policy*. 781
- 747 Luoga, E., Witkowski, E.T.F., Balkwill, K., 2002. Harvesting and
748 standing wood stocks in protected and communal miombo
749 woodlands of eastern Tanzania. *Forest Ecology and
750 Management* 164, 15–30. 782
- 751 Lwanga, J.S., 2003. Forest succession in Kibale National Park,
752 Uganda: Implications for forest restoration and management.
753 *African Journal of Ecology* 41, 9–22. 783
- 754 Mooney, H.A., Godron, M., 1983. *Disturbance and Ecosystems:
755 Components of Response*. Springer-Verlag, New York. 784
- 756 Mugisha, R.A., 2002. Evaluation of community-based
757 conservation approaches: management of protected areas in
758 Uganda. Ph.D. dissertation, University of Florida, Gainesville,
759 Florida. 785
- 760 Mulley, B., Unruh, J., 2004. The role of off-farm employment in
761 tropical forest conservation: labor, migration, and smallholder
762 attitudes towards land in western Uganda. *Journal of
763 Environment Management* 71, 193–205. 786
- 764 Naughton-Treves, L., 1998. Predicting patterns of crop damage by
765 wildlife around Kibale National Park, Uganda. *Conservation
766 Biology* 12, 156–168. 787
- 767 Naughton-Treves, L., 1999. Whose Animals? A history of property
768 rights to wildlife in Toro, western Uganda. *Land Degradation
769 and Development* 10, 311–328. 788
- 770 Naughton-Treves, L., Chapman, C., 2002. Fuelwood resources on
771 fallow land in East Africa. *Journal of Sustainable Forestry* 14,
772 19–32. 789
- 773 Okello, B.D., O'Connor, T.G., Young, T.P., 2001. Growth, biomass
774 estimates, and charcoal production of *Acacia drepanolobium*
in Laikipia, Kenya. *Forest Ecology and Management* 142, 143–
153. 790
- Parikka, M., 2004. Global biomass fuel resources. *Biomass and
Bioenergy* 27, 613–620. 791
- Parrotta, J.A., Francis, J.K., Knowles, O.H., 2002. Harvesting
intensity affects forest structure and composition in an
upland Amazonian forest. *Forest Ecology and Management*
169, 243–255. 792
- Paul, J.R., Randle, A.M., Chapman, C.A., Chapman, L.J., 2004.
Arrested succession in logging gaps: is tree seedling growth
and survival limiting. *African Journal of Ecology* 42, 245–251. 793
- Plumptre, A., 2002. Extent and status of forests in the Albertine
Rift, *Wildlife Conservation Society Albertine Rift Program*,
Bronx, New York. 794
- Ribot, J., 1998. Theorizing access: forest profits along Senegal's
charcoal commodity chain. *Development and Change* 29, 307–
341. 795
- Rocheleau, D., Edmunds, D., 1997. Women, men and trees: gender,
power and property in forest and agrarian landscapes. *World
Development* 25, 1351–1371. 796
- Silvertown, J.W., Lovett-Doust, J., 1993. *Introduction to Plant
Population Biology*. Blackwell Scientific Publications,
Cambridge, Great Britain. 797
- Skorupa, J., 1988. The effect of selective timber harvesting on rain-
forest primates in Kibale Forest, Uganda, Ph.D. Dissertation. 798
- Struhsaker, T.T., 1987. Forestry issues and conservation in
Uganda. *Biological Conservation* 39, 209–234. 799
- Struhsaker, T.T., 1997. *Ecology of an African Rain Forest*.
University Press of Florida, Gainesville, FL. 800
- Struhsaker, T.T., Struhsaker, P.J., Siex, K.S., 2005. Conserving
Africa's rain forests: problems in protected areas and possible
solutions. *Biological Conservation* 123, 45–54. 801
- Turner, G.M., Romme, W.H., Gardener, R.H., O'Neill, R.V., Krtz, T.K.,
1993. A revised concept of landscape equilibrium: disturbance
and stability of on scaled landscapes. *Landscape Ecology* 8,
213–227. 802
- White, P.S., Pickett, S.T.A., 1985. Natural disturbance and patch
dynamics: an introduction. In: Pickett, S.T.A., White, P.S. (Eds.),
The Ecology of Natural Disturbance and Patch Dynamics.
Wiley-Interscience, New York, TJ, pp. 1–13. 803
- World Bank, 1993. *Uganda. Agriculture*. The World Bank,
Washington, DC. 804
- Zanne, A.E., Chapman, C.A., in press. Diversity of woody species
in forest, tree fall gaps, and edge in Kibale National Park,
Uganda, *Plant Ecology*. 805